

## Some theoretical results concerning O<sub>3</sub>-NO<sub>x</sub>-VOC chemistry and NO<sub>x</sub>-VOC indicators

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[1] A series of model results is shown pertaining to ozone, reactive nitrogen (NO<sub>y</sub>), and peroxides in polluted regions. The results focus on ratios such as O<sub>3</sub>/NO<sub>y</sub> and H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> that have been proposed as indicators for O<sub>3</sub>-NO<sub>x</sub>-VOC sensitivity. These ratios are shown to correlate with predicted NO<sub>x</sub>-VOC sensitivity for a variety of zero-dimensional (0-D) and 3-D models, but the correlation varies in situations ranging from relatively clean to highly polluted. The previously identified NO<sub>x</sub>-VOC transition values for indicators appear to be valid for moderately polluted conditions with 80–150 ppb O<sub>3</sub>. Changes in indicator behavior are linked to odd hydrogen chemistry and in particular to changes in the ratio of O<sub>3</sub> to primary radical production. Changes in indicator behavior are also correlated with the values of a test ratio, O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub>+NO<sub>z</sub>) which can be evaluated against measurements. Ratios of the form ΔO<sub>3</sub>/ΔNO<sub>y</sub>, representing differences relative to background values, are proposed for analyzing NO<sub>x</sub>-VOC sensitivity in individual urban plumes. Comparisons are made to extent-of-reaction parameters, which have been proposed for evaluating NO<sub>x</sub>-VOC sensitivity. *INDEX TERMS*: 0365 Atmospheric Composition and Structure: Troposphere—composition and chemistry; 0345 Atmospheric Composition and Structure: Pollution—urban and regional (0305); 3210 Mathematical Geophysics: Modeling; *KEYWORDS*: ozone, nitrogen oxides, volatile organic compounds (VOC), hydrogen peroxide, photochemical smog

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### 1. Introduction

[2] The chemistry of ozone and its two main precursors, NO<sub>x</sub> and volatile organic compounds (VOC), continues to represent one of the major uncertainties in the field of atmospheric chemistry. In urban areas, uncertainty associated with O<sub>3</sub>-NO<sub>x</sub>-VOC chemistry can affect the design of control strategies to reduce ambient O<sub>3</sub>. In the remote troposphere, O<sub>3</sub>-NO<sub>x</sub>-VOC chemistry affects evaluations of the ozone production efficiency per NO<sub>x</sub> as well as predicted responses to future changes in emissions.

[3] In recent years a number of works have analyzed the photochemical factors that determine the split into VOC-sensitive and NO<sub>x</sub>-sensitive regimes [Sillman *et al.*, 1990; Sillman, 1995; Kleinman, 1994; Kleinman *et al.*, 1997; Tonnesen and Dennis, 2000a, 2000b; Jaegle *et al.*, 1998, 2001]. Sillman [1995] and Sillman *et al.* [1998] proposed that O<sub>3</sub>-NO<sub>x</sub>-VOC chemistry could be linked to the ratios of certain measurable species which had different values for NO<sub>x</sub>-sensitive and VOC-sensitive conditions. If successful, these “NO<sub>x</sub>-VOC indicators” can provide a powerful tool for evaluating the chemical process leading to ozone formation. However, there is a range of theoretical and practical concerns associated with the proposed indicators. Contradictory results were reported by Lu and

Chang [1998] and Chock *et al.* [1999]. It is especially unclear whether the indicator ratios would show similar behavior for a wide variety of conditions, as proposed by Sillman.

[4] Here, a series of model calculations is used to explore indicator behavior for a variety of conditions, ranging from rural to extremely polluted conditions. The calculations have two purposes: to identify how the behavior of indicator ratios may vary under different conditions; and to identify the link between the proposed indicator ratios and the chemistry of odd hydrogen radicals. The link between indicator ratios and odd hydrogen radicals, presented previously by Sillman [1995], Kleinman [1994], and Kleinman *et al.* [1997] will be used to explain variations in the behavior of indicator ratios for different conditions. Model results will also be used to identify broad correlation patterns among ozone, reactive nitrogen, and peroxides that are associated with NO<sub>x</sub>-sensitive and VOC-sensitive chemistry.

[5] Results will be shown from both zero-dimensional (0-D) calculations and 3-D models. It should be emphasized that valid predictions for species concentrations must be based on 3-D models, which include more detailed representation of dynamical processes and which include comparison with measured values. The 0-D calculations are useful because they can extend previous analyses to new conditions (which do not match observed cases, but which may correspond to future situations) and because they allow

a more detailed analysis of the chemical factors which drive the results.

## 2. Theoretical Background

[6] The split between NO<sub>x</sub>-sensitive and VOC-sensitive conditions is well-known, and illustrated by ozone isopleth plots (e.g., see Figure 1). For conditions with relatively high VOC and low NO<sub>x</sub>, O<sub>3</sub> increases with increasing NO<sub>x</sub> and is relatively insensitive to changes in VOC. For conditions with relatively low VOC and high NO<sub>x</sub>, O<sub>3</sub> increases with increasing VOC and decreases with increasing NO<sub>x</sub>. An analogous split between “NO<sub>x</sub>-sensitive” and “NO<sub>x</sub>-saturated” regimes occurs in the remote troposphere, although for remote conditions O<sub>3</sub> increases with increasing VOC even in the NO<sub>x</sub>-sensitive regime [Jaegle *et al.*, 1998, 2001].

[7] The split between NO<sub>x</sub>-sensitive and VOC-sensitive regimes is driven by the chemistry of odd hydrogen radicals. As shown by Sillman *et al.* [1990] and Sillman [1995], the NO<sub>x</sub>-VOC split was attributed to the relative rate of formation of peroxides (via HO<sub>2</sub>-HO<sub>2</sub> and HO<sub>2</sub>-RO<sub>2</sub> reactions) relative to nitric acid formation (via OH + NO<sub>2</sub>). NO<sub>x</sub>-sensitive conditions occur when peroxides dominate over nitric acid as radical sinks, while NO<sub>x</sub>-saturated conditions occur when nitric acid dominates. As shown by Kleinman [1994], NO<sub>x</sub>-VOC sensitivity was attributed to the relative source strengths of odd hydrogen radicals (S<sub>H</sub>) and odd nitrogen (S<sub>N</sub>), summed over the period of ozone production for an air parcel. (Abbreviations are summarized in section 8.) VOC-sensitive chemistry occurred when the odd nitrogen source exceeded the source of odd hydrogen. Subsequently Kleinman *et al.* [1997] found that the instantaneous rate of ozone production was VOC-sensitive whenever the instantaneous loss rate for odd nitrogen (L<sub>N</sub>) was greater than half the total odd hydrogen source (L<sub>N</sub>/Q > 0.5, as given by Kleinman *et al.* [1997]). This formulation is equivalent to that of Sillman [1995] if net formation of PAN and other organic nitrates is assumed to be zero. From Sillman [1995], odd hydrogen sources must be in steady state with its three major sinks:

$$S_H = 2P_{\text{perox}} + P_{\text{HNO}_3} + P_{\text{pans}} \quad (1)$$

where P<sub>perox</sub> and P<sub>HNO<sub>3</sub></sub> represent production rates for peroxides (including H<sub>2</sub>O<sub>2</sub> and organic peroxides) and HNO<sub>3</sub>, and P<sub>pans</sub> represents net photochemical production of PAN and higher order analogues. VOC-sensitive chemistry would occur whenever P<sub>HNO<sub>3</sub></sub> exceeds 2P<sub>perox</sub>. In terms of S<sub>H</sub>, VOC-sensitive chemistry would occur for the following conditions:

$$S_H < \xi L_N \quad (2a)$$

$$\xi = 1 + \frac{P_{\text{HNO}_3}}{P_{\text{HNO}_3} + P_{\text{PANs}}} \quad (2b)$$

or

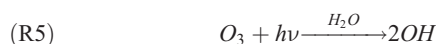
$$S_H - P_{\text{pans}} < 2P_{\text{HNO}_3} \quad (2c)$$

[8] This modified form serves to emphasize that the NO<sub>x</sub>-VOC split is associated with formation of nitric acid but is not affected by formation of PAN [Sillman *et al.*, 1990; Tonnesen and Dennis, 2000b].

[9] The above results were derived by Sillman [1995] and Kleinman *et al.* [1997] based on the following simplified photochemistry. The ozone production sequence includes



[10] Similar reaction sequences apply for many individual VOC, which produce various radical chains with the form RO<sub>2</sub> and with subsequent reactions analogous to (R2). Reaction (R1) (VOC + OH) is the rate limiting step for this sequence, and its rate depends on the availability of OH. OH in turn depends on the balance of sources and sinks of odd hydrogen radicals (including OH, HO<sub>2</sub> and RO<sub>2</sub>), which includes the following sources

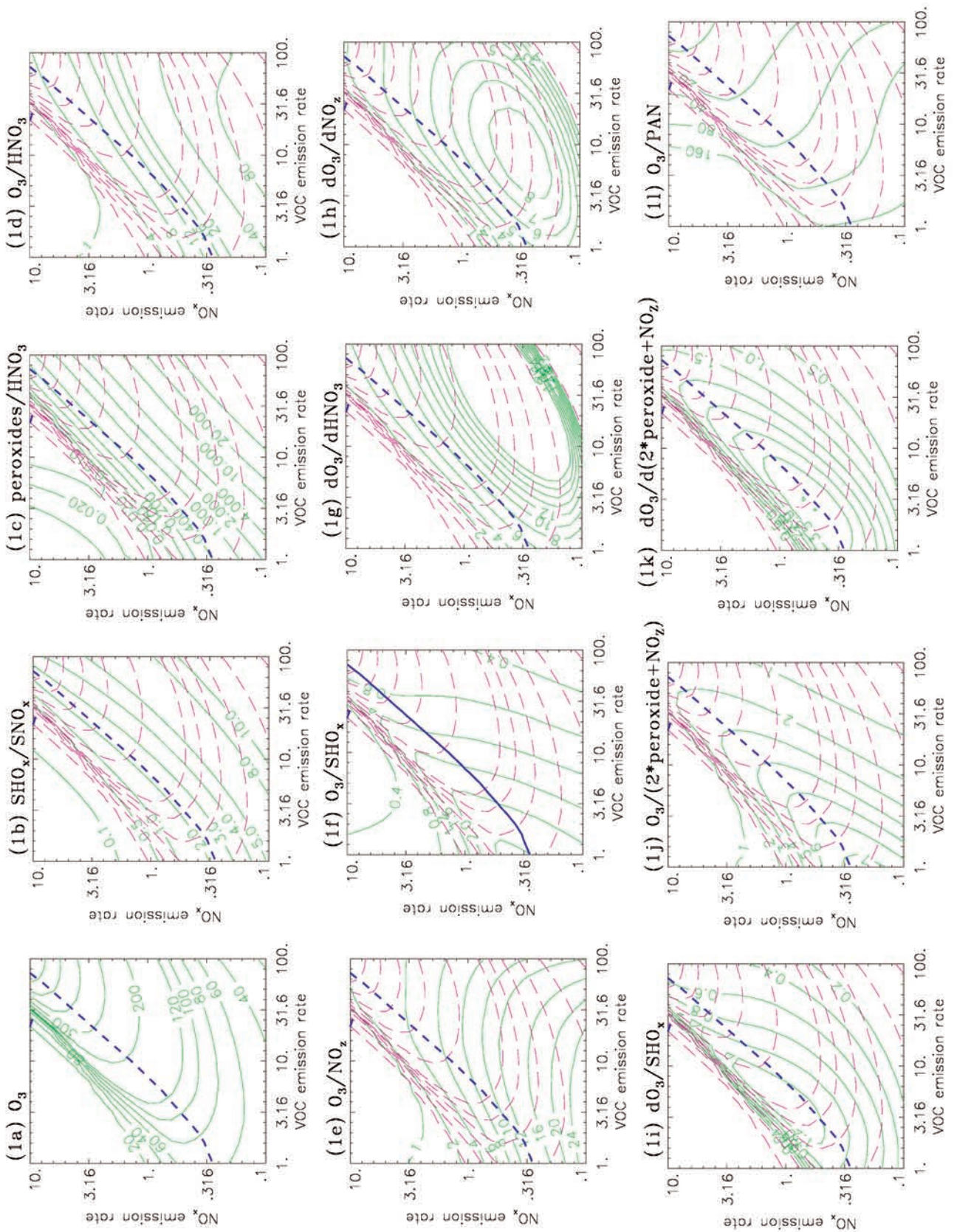


[11] Other sources include the photolysis of higher aldehydes and various alkene-O<sub>3</sub> reactions [Paulson and Orlando, 1996]. Sinks include the following.



where (R10) refers to net production of PAN. As first developed by Sillman *et al.* [1990], when the peroxide-

**Figure 1.** (opposite) Isopleths as a function of the average emission rate for NO<sub>x</sub> and VOC (10<sup>12</sup> molec. cm<sup>-2</sup> s<sup>-1</sup>) in 0-D calculations. The isopleths represent conditions in the first layer during the afternoon following 3-day calculations, at the hour corresponding to maximum O<sub>3</sub>. Isopleths are shown for (a) O<sub>3</sub> (ppb); (b) S<sub>H</sub>/S<sub>N</sub>; (c) (H<sub>2</sub>O<sub>2</sub> + ROOH)/HNO<sub>3</sub>; (d) O<sub>3</sub>/HNO<sub>3</sub>; (e) O<sub>3</sub>/NO<sub>z</sub>; (f) O<sub>3</sub>/S<sub>H</sub>; (g) ΔO<sub>3</sub>/ΔHNO<sub>3</sub>; (h) ΔO<sub>3</sub>/ΔNO<sub>z</sub>; (i) ΔO<sub>3</sub>/S<sub>H</sub>; (j) O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>); (k) ΔO<sub>3</sub>/Δ(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>); and (l) O<sub>3</sub>/PAN. Isopleths are shown as solid green lines. Isopleths for O<sub>3</sub> (red dashed lines) are superimposed on the other isopleth plots. The short blue dashed line represents the transition from VOC-sensitive to NO<sub>x</sub>-sensitive conditions.



forming reactions (R7) and (R8) are the dominant radical sinks, then the radical steady state reduces to

$$HO_2 \propto \sqrt{S_H - P_{pans}} \quad (3)$$

where the similar but more complex RO<sub>2</sub> has been omitted for simplicity. The production rate for O<sub>3</sub> is approximately equal to the rate of reactions (R2) and (R3) (HO<sub>2</sub> + NO, RO<sub>2</sub> + NO). This rate increases with increasing NO<sub>x</sub> but has little direct dependence on VOC. By contrast, when the reaction to form nitric acid (R9) is the dominant radical sink, then the radical steady state becomes

$$OH \propto (S_H - P_{pans})/NO_2 \quad (4)$$

[12] The production rate for O<sub>3</sub> is proportional to the rate of the VOC + OH reactions (R1), and this rate increases with increasing VOC and decreases with increasing NO<sub>x</sub>.

[13] *Tonnesen and Dennis* [2000a] presented a similar results with slightly different terminology. Tonnesen and Dennis analyzed ozone formation in terms of radical formation (equivalent to S<sub>H</sub> here), radical termination (through production of peroxides, HNO<sub>3</sub> and organic nitrates) and radical propagation (through VOC + OH and HO<sub>2</sub> + NO, reactions (R1), (R2), and (R3)). The radical propagation reactions are directly associated with ozone formation and were regarded as synonymous with ozone production in *Sillman* [1995].

[14] *Kleinman et al.* [1997], *Tonnesen and Dennis* [2000a], and *Kirchner et al.* [2001] have developed methods for evaluating whether the instantaneous rate of ozone production is sensitive to NO<sub>x</sub> or VOC. This instantaneous sensitivity should be distinguished from the sensitivity associated with ozone concentrations, which are affected by upwind emissions, transport, and photochemistry. As discussed by *Kirchner et al.* [2001], *Sillman* [1995] identified indicator ratios that relate specifically to the NO<sub>x</sub>-VOC sensitivity of ozone concentrations rather than instantaneous production rates. The subsequent results all relate to the NO<sub>x</sub>-VOC dependence of ozone concentrations, based on photochemical production over extended periods of time.

[15] The NO<sub>x</sub>-VOC indicators proposed by *Sillman* [1995] included the following: O<sub>3</sub>/NO<sub>y</sub> (where NO<sub>y</sub> represents total reactive nitrogen); O<sub>3</sub>/NO<sub>z</sub> (where NO<sub>z</sub> represents summed NO<sub>x</sub> reaction products, or NO<sub>y</sub>-NO<sub>x</sub>); O<sub>3</sub>/HNO<sub>3</sub>; H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub>; H<sub>2</sub>O<sub>2</sub>/NO<sub>z</sub>; and the equivalent ratios with summed H<sub>2</sub>O<sub>2</sub> and organic peroxides. The ratios involving peroxides were justified based on the role of peroxides and nitric acid as sinks for odd hydrogen radicals. It is more difficult to explain why ratios such as O<sub>3</sub>/NO<sub>z</sub> should be associated with NO<sub>x</sub>-VOC sensitivity. *Sillman* [1995] suggested that O<sub>3</sub> was roughly proportional to the odd hydrogen source, S<sub>H</sub>. The ratio O<sub>3</sub>/NO<sub>z</sub> is then analogous to S<sub>H</sub>/L<sub>N</sub>, which is related to NO<sub>x</sub>-VOC sensitivity for reasons given above. This will be discussed in more detail in section 4.

### 3. Methods

[16] Results will be shown from a series of simplified 0-D calculations and from previously published 3-D

model simulations. The 0-D calculations use 2-layer model [*Sillman et al.*, 1990] which has been adapted from urban photochemical box models. The model consists of a lower layer, representing the ambient boundary layer, and an upper residual layer representing conditions aloft at night. During the morning hours the height of the lower layer expands and the contents of the residual layer become entrained into the model lower layer. During the evening hours the height of the lower layer decreases and the remainder becomes part of the residual layer. The height of the lower layer varies from 200 m. at night to 1500 m. in the afternoon, representing typical behavior for a boundary layer during pollution events. The model residual layer extends from the top of the lower layer up to 1600 m. Emissions are entered into the lower layer as concentrations based on specified emission rates and the layer height. Although this model includes some rudimentary dynamics in one (vertical) dimension, it will be referred to as a 0-D model, as realistic dynamics are not included.

[17] Photochemistry and dry deposition are as shown by *Sillman et al.* [1998]. The photochemical mechanism is based on the work of *Lurmann et al.* [1986] with various modifications, including updated reaction rates from *DeMore et al.* [1997], isoprene chemistry from *Paulson and Seinfeld* [1992], RO<sub>2</sub>-RO<sub>2</sub> reactions from *Kirchner and Stockwell* [1996], and organic peroxide reaction rates from *Stockwell et al.* [1997]. Aerosol reactions are not included. One weakness of this photochemical representation is that it has not been tested against environmental chamber experiments. By contrast, the RACM [*Stockwell et al.*, 1997] and SAPRC [*Carter*, 2000] have been extensively tested against chamber data. The current mechanism has been retained here because it includes many reactions that are important in the rural/remote troposphere. Also, because aerosol reactions are not included, the calculations do not account for the formation of aerosol nitrate (NO<sub>3</sub><sup>-</sup>) from heterogeneous reaction with ammonia. HNO<sub>3</sub> in these calculations should be viewed as equivalent to the sum of HNO<sub>3</sub> and NO<sub>3</sub><sup>-</sup> [*Martilli et al.*, 2002].

[18] Calculations were performed for 3-day time periods with diurnally varying emissions, for a wide variety of anthropogenic VOC and NO<sub>x</sub> emission rates. Speciation of VOC was based on average speciation in the NAPAP 1990 inventory [*Environmental Protection Agency (EPA)*, 1993]. Diurnal variations were based on average diurnal variations in the same inventory. In the standard series of calculations emission rates were constant (except for the diurnal variation) throughout the 3-day period. Alternative calculations were performed with higher emissions on the third day, representing a situation in which processed rural air would enter an urban area on the third day. Initial conditions were typical for remote air in the U.S.: 40 ppb O<sub>3</sub>, 1 ppb H<sub>2</sub>O<sub>2</sub>, 15 ppt NO and NO<sub>2</sub>, and 5 ppb C VOC.

[19] The 3-D simulations are summarized in Table 1.

### 4. Results From 0-D Calculations

[20] Figure 1 shows isopleths for O<sub>3</sub> and for a series of species ratios, shown as a function of diurnal average emission rates for VOC and NO<sub>x</sub>. The plots represent

**Table 1.** 3-D Simulations Used in This Study<sup>a</sup>

| Location            | Model                                    | Photochemistry                           | Model Domain   | Comparison With Measurements                                 | Reference  |
|---------------------|--|--|--|--|--|
| Nashville*          | <i>Sillman et al.</i> [1998]             | modified [ <i>Lurmann et al.</i> , 1986] | 5 × 5 km urban; upwind domains includes eastern U.S.       | O <sub>3</sub> , NO <sub>y</sub> , peroxides                 | <i>Sillman et al.</i> [1998]                                       |
| Lake Michigan*      | <i>Sillman et al.</i> [1993]             | modified [ <i>Lurmann et al.</i> , 1986] | 20 × 20 km in region; upwind domains includes eastern U.S. | O <sub>3</sub>   | <i>Sillman</i> [1995]  |
| Northeast corridor* | <i>Sillman et al.</i> [1993]             | modified [ <i>Lurmann et al.</i> , 1986] | 20 × 20 km in region; upwind domains includes eastern U.S. | O <sub>3</sub>   | <i>Sillman</i> [1995]  |
| Atlanta             | UAM-IV [ <i>Morris and Myers</i> , 1990] | CB4 [ <i>Gery et al.</i> , 1989]         | 5 × 5 km urban   | O <sub>3</sub> , NO <sub>y</sub> , isoprene, HCHO, other VOC | <i>Sillman et al.</i> [1997]                                       |
| San Joaquin*        | MAQSIP [ <i>Odman and Ingram</i> , 1996] | CB4 [ <i>Gery et al.</i> , 1989]         | 12 × 12 km; domain includes all central California         | O <sub>3</sub>   | <i>Sillman et al.</i> [2001]                                       |
| San Joaquin         | SAQM [ <i>J. Chang et al.</i> , 1997]    | CB4 [ <i>Gery et al.</i> , 1989]         | 12 × 12 km; domain includes all central California         | O <sub>3</sub>   | <i>Lu and Chang</i> [1998]   |
| Los Angeles*        | UAM-IV [ <i>Morris and Myers</i> , 1990] | CB4 [ <i>Gery et al.</i> , 1989]         | 5 × 5 km urban   | O <sub>3</sub> , NO <sub>y</sub> , NO <sub>2</sub>           | <i>Godowitch and Vukovich</i> [1994]; <i>Sillman et al.</i> [1997] |
| Los Angeles         | UAM-IV [ <i>Morris and Myers</i> , 1990] | CB4 [ <i>Gery et al.</i> , 1989]         | 5 × 5 km urban   | O <sub>3</sub>   | <i>Chock et al.</i> [1999]   |

<sup>a</sup> Asterisks (\*) denote models included in Figures 3–6.

conditions during the afternoon of the third day of a 3-day calculation, at the hour corresponding to maximum O<sub>3</sub>. Figure 1a is the standard isopleth plot for O<sub>3</sub> and illustrates the well-known split between NO<sub>x</sub>-sensitive and VOC-sensitive (or NO<sub>x</sub>-saturated) conditions. Figure 1a also shows the “sensitivity transition” from NO<sub>x</sub>-sensitive to VOC-sensitive conditions, defined as follows:

$$\frac{1}{Q_N} \frac{\partial [O_3]}{\partial Q_N} = \frac{1}{Q_H} \frac{\partial [O_3]}{\partial Q_H} \quad (5)$$

where [O<sub>3</sub>] represents ozone concentrations and Q<sub>N</sub> and Q<sub>H</sub> represent emission rates for NO<sub>x</sub> and VOC. Using this definition, the sensitivity transition represents the point at which a given percent reduction in either NO<sub>x</sub> or VOC would result in the same reduction in O<sub>3</sub>. As shown in Figure 1a, reduced NO<sub>x</sub> would result in lower O<sub>3</sub> than an equivalent percent reduction in VOC when Q<sub>H</sub>/Q<sub>N</sub> is higher than its value at the transition line. When Q<sub>H</sub>/Q<sub>N</sub> is lower than the transition value, then reduced VOC would result in lower O<sub>3</sub> than an equivalent percent reduction in NO<sub>x</sub>. The same transition line is included for reference in the other isopleth plots.

[21] Figure 1b shows the calculated ratio S<sub>H</sub>/S<sub>N</sub>, where S<sub>H</sub> and S<sub>N</sub> represent calculated sums of the source of odd hydrogen and odd nitrogen over the 3-day period. This ratio is similar to the ratio used by *Kleinman et al.* [1997] to analyze instantaneous ozone chemistry, but here they represent sums for an extended period of ozone production (including nighttime) rather than an instantaneous conditions. There is some ambiguity about this ratio in multiday calculations, because the exact value of S<sub>H</sub>/S<sub>N</sub> depend on the time period over which S<sub>H</sub> is summed and because for realistic conditions an air parcel will include a wide variety of ozone production rates over its photochemical history. In

these simplified calculations, ozone production occurs continuously over the 3-day period with relatively little variation for each individual calculation. We have summed S<sub>H</sub> and S<sub>N</sub> over the full 3 days for both model layers as molecules produced or emitted (rather than molecules per volume). As shown below, this provides a useful way to identify differences between calculations that represent equivalent air mass histories but with very different amounts of precursors.

[22] As shown in the figure, the ratio S<sub>H</sub>/S<sub>N</sub> is closely associated with NO<sub>x</sub>-VOC sensitivity and the transition from NO<sub>x</sub>-sensitive to VOC-sensitive conditions is associated with a specific value of S<sub>H</sub>/S<sub>N</sub>. This transition value (S<sub>H</sub>/S<sub>N</sub> = 2) remains for emission rates varying by a factor of 100, and for O<sub>3</sub> varying from 60 to 600 ppb. Higher S<sub>H</sub>/S<sub>N</sub> is associated with NO<sub>x</sub>-sensitive conditions and lower values with VOC-sensitive conditions. An equivalent transition value (L<sub>N</sub>/S<sub>H</sub> = 0.5) was found by *Kleinman et al.* [1997] for the NO<sub>x</sub>-VOC variation of instantaneous ozone production. The ratio S<sub>H</sub>/S<sub>N</sub> cannot be measured, but it provides a theoretical basis for identifying measurable species ratios as potential NO<sub>x</sub>-VOC indicators.

[23] The region with S<sub>H</sub>/S<sub>N</sub> < 1 is also of interest, because this corresponds to conditions in which the source of radicals is not sufficient to oxidize the total amount of NO<sub>x</sub> emitted into the air mass. In this region ozone concentrations are sharply lower. The ratio NO<sub>x</sub>/NO<sub>3</sub>, a measure of the extent of photochemical processing, increases sharply for S<sub>H</sub>/S<sub>N</sub> less than one (see Figure 7). This is discussed more in section 6.

[24] Figure 1c shows the ratio (H<sub>2</sub>O<sub>2</sub> + ROOH)/HNO<sub>3</sub>. The value of this ratio is also closely associated with NO<sub>x</sub>-VOC sensitivity, although the NO<sub>x</sub>-VOC transition value is not uniformly constant. The transition value is 0.5 for O<sub>3</sub> below 250 ppb. A slightly higher value (0.5–0.6) was found in the 3-D model reported by *Sillman et al.* [1998].

At higher emission rates and higher O<sub>3</sub> the transition value drifts lower, reaching 0.2 at 500 ppb O<sub>3</sub>. The NO<sub>x</sub>-sensitive and VOC-sensitive regions of the isopleth plot are associated respectively with much higher and lower values of the ratio.

[25] The ratio H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> shows a similar close association with NO<sub>x</sub>-VOC sensitivity, and again the NO<sub>x</sub>-VOC transition value tends to decrease at higher O<sub>3</sub> and higher emission rates. Referring to Figure 1c, the most useful information about these ratios is the value along the sensitivity transition of the isopleth plot. These values are shown plotted against O<sub>3</sub> along the sensitivity transition in Figure 2. The transition value for H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> varies from 0.3 at 100 ppb O<sub>3</sub> to 0.07 at 400 ppb O<sub>3</sub>. The transition values at 150 ppb O<sub>3</sub> in these calculations all correspond closely with transition values in 3-D simulations reported by *Sillman et al.* [1998].

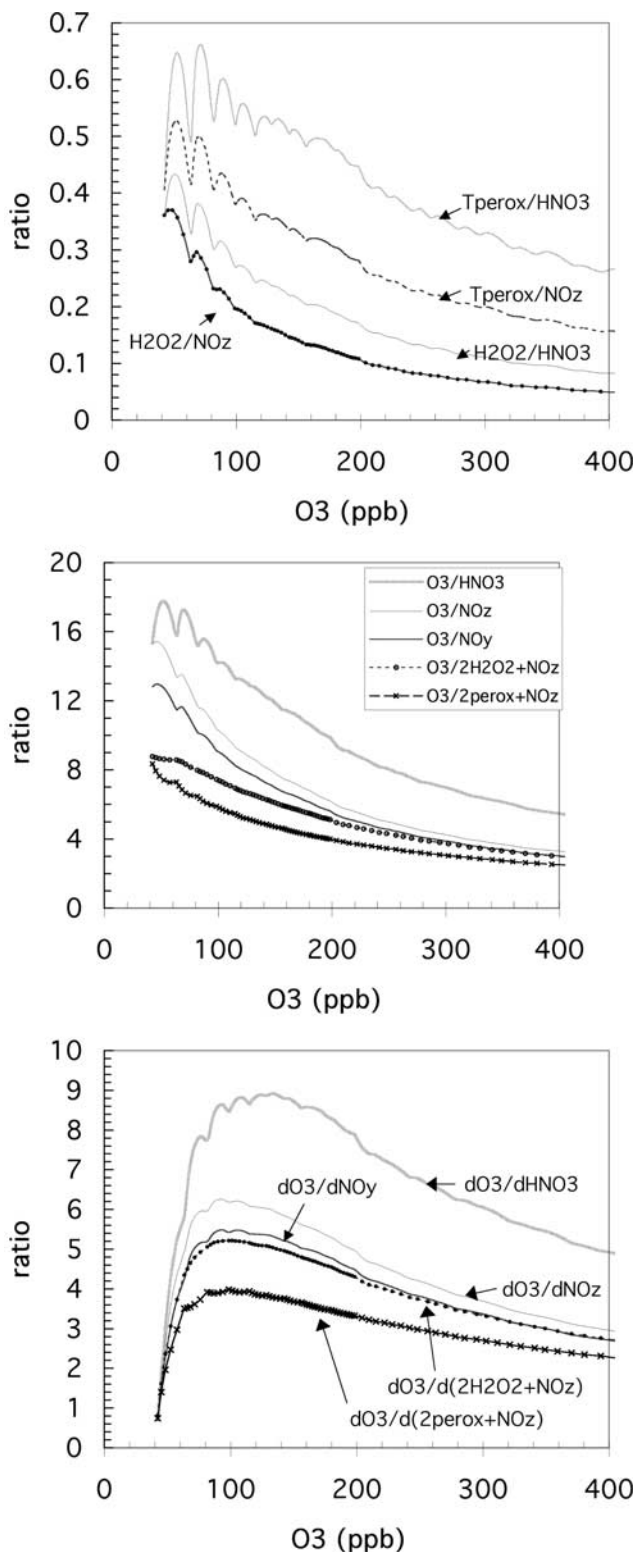
[26] The ratio O<sub>3</sub>/HNO<sub>3</sub> (Figure 1d) both show a general correspondence with NO<sub>x</sub>-VOC sensitivity, but the correspondence is not as clear and the transition value shows greater variation. The ratio O<sub>3</sub>/NO<sub>z</sub> (Figure 1e) shows a significantly worse correspondence with NO<sub>x</sub>-VOC sensitivity. There is a general (though imperfect) correspondence between O<sub>3</sub>/NO<sub>z</sub> and NO<sub>x</sub>-VOC sensitivity for O<sub>3</sub> below 200 ppb, but the correspondence breaks down for higher O<sub>3</sub>. When VOC emissions are high (corresponding to high O<sub>3</sub> at the sensitivity transition) the ratio O<sub>3</sub>/NO<sub>z</sub> appears to vary in direct proportion to VOC emissions and does not increase (or may even decrease) with increasing NO<sub>x</sub>. The transition value of O<sub>3</sub>/HNO<sub>3</sub> and O<sub>3</sub>/NO<sub>z</sub> (i.e., the value along the NO<sub>x</sub>-VOC sensitivity transition) decreases as emission rates increase. This variation along the transition value is also shown in Figure 2b. Similar results were reported by *Tonnesen and Dennis* [2000b] and *Kirchner et al.* [2001].

[27] The behavior of O<sub>3</sub>/NO<sub>z</sub> and O<sub>3</sub>/HNO<sub>3</sub> can be understood by regarding them as an imperfect surrogate for Kleinman's ratios S<sub>H</sub>/S<sub>N</sub> or S<sub>H</sub>/L<sub>N</sub>. It was shown above that S<sub>H</sub>/S<sub>N</sub> is correlated with NO<sub>x</sub>-VOC sensitivity for the entire range of 0-D calculations used here. O<sub>3</sub>/NO<sub>z</sub> can be expressed as a function of S<sub>H</sub>/L<sub>N</sub> as follows:

$$\frac{O_3}{NO_z} = \left[ \frac{S_H}{L_N} \right] \left[ \frac{L_N}{NO_z} \right] \left[ \frac{O_3}{S_H} \right] \quad (6a)$$

[28] Based on Figure 1b, if O<sub>3</sub>/NO<sub>z</sub> were directly proportional to S<sub>H</sub>/S<sub>N</sub> it would be expected to correlate with NO<sub>x</sub>-VOC sensitivity and show a near-constant value along the sensitivity transition throughout the range of model calculations. To the extent that O<sub>3</sub>/NO<sub>z</sub> shows a different pattern, the differences should be associated with the terms L<sub>N</sub>/NO<sub>z</sub> and O<sub>3</sub>/S<sub>H</sub> in equation (6a). Therefore, L<sub>N</sub>/NO<sub>z</sub> and O<sub>3</sub>/S<sub>H</sub> may be regarded as sources of error for O<sub>3</sub>/NO<sub>z</sub> as a NO<sub>x</sub>-VOC indicator.

[29] The ratio L<sub>N</sub>/NO<sub>z</sub> relates the rate of removal of NO<sub>x</sub> (primarily through conversion to the species included in NO<sub>z</sub>) to the NO<sub>z</sub> concentration. As a source of error, this relates to the removal rate of NO<sub>z</sub>. The ratio O<sub>3</sub>/S<sub>H</sub> is more important as a source of error in the indicator concept. *Sillman* [1995] assumed that S<sub>H</sub> was likely to be proportional to O<sub>3</sub> because O<sub>3</sub> is a major precursor of odd hydro-



**Figure 2.** Values of ratios along the NO<sub>x</sub>-VOC transition line shown in Figure 1, for 0-D calculations. Values are shown for (a) (H<sub>2</sub>O<sub>2</sub> + ROOH)/HNO<sub>3</sub>, (H<sub>2</sub>O<sub>2</sub> + ROOH)/NO<sub>z</sub>, H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub>, and H<sub>2</sub>O<sub>2</sub>/NO<sub>z</sub>; (b) O<sub>3</sub>/HNO<sub>3</sub>, O<sub>3</sub>/NO<sub>z</sub>, O<sub>3</sub>/NO<sub>y</sub>, O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>), and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) (listed in order from highest to lowest); and (c) ΔO<sub>3</sub>/ΔHNO<sub>3</sub>, ΔO<sub>3</sub>/ΔNO<sub>z</sub>, ΔO<sub>3</sub>/ΔNO<sub>y</sub>, ΔO<sub>3</sub>/Δ(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>), and ΔO<sub>3</sub>/Δ(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>). The sum H<sub>2</sub>O<sub>2</sub> + ROOH is abbreviated as “perox” in the figure labels.

gen radicals. However, VOC are also important radical sources, so that O<sub>3</sub>/S<sub>H</sub> can be expected to decrease with increasing VOC. O<sub>3</sub>/S<sub>H</sub> can also vary photochemical age, since S<sub>H</sub> here represents the summed source over the period of ozone production. The ratio O<sub>3</sub>/NO<sub>y</sub> (included in Figure 2b) is useful in this regard because it indirectly accounts for photochemical aging. O<sub>3</sub>/NO<sub>y</sub> can be viewed as O<sub>3</sub>/NO<sub>z</sub> multiplied by an aging term (NO<sub>z</sub>/NO<sub>y</sub>).

[30] As shown in Figure 1f, the ratio O<sub>3</sub>/S<sub>H</sub> decreases with increasing VOC emissions along the sensitivity transition. The change in O<sub>3</sub>/S<sub>H</sub> is proportional to the change in O<sub>3</sub>/NO<sub>z</sub> along the sensitivity transition, and thus explains much of the change in the transition value of O<sub>3</sub>/NO<sub>z</sub> from low to high emissions.

[31] As an alternative approach, it is possible to use difference ratios such as ΔO<sub>3</sub>/ΔNO<sub>z</sub>, where ΔO<sub>3</sub> and ΔNO<sub>z</sub> represent differences relative to background values ((O<sub>3</sub>-O<sub>3b</sub>) and (NO<sub>z</sub>-NO<sub>zb</sub>)). The difference ratio is related to the O<sub>3</sub>-NO<sub>z</sub> slope, which has been used to evaluate the ozone production efficiency per NO<sub>x</sub> [Trainer *et al.*, 1993]. The difference ratio is related to Kleinman's S<sub>H</sub>/L<sub>N</sub> as follows:

$$\frac{\Delta O_3}{\Delta NO_z} = \left[ \frac{S_H}{L_N} \right] \left[ \frac{L_N}{\Delta NO_z} \right] \left[ \frac{\Delta O_3}{S_H} \right] \quad (6b)$$

[32] The error term is now ΔO<sub>3</sub>/S<sub>H</sub>. This term can be interpreted as an ozone production efficiency per primary radical production [see Daum *et al.*, 2000a]. However, since the reaction sequence leading to production of radicals from secondary hydrocarbons is similar to the reaction sequence leading to production of O<sub>3</sub>, it is possible to view ΔO<sub>3</sub>/S<sub>H</sub> as a ratio of two closely related photochemical processes. As such, this ratio is likely to show little variation over a wide range of photochemical conditions. The ratio ΔO<sub>3</sub>/S<sub>H</sub> is also related to the "chain length", the ratio of chain propagation to chain termination described by Tonnesen and Dennis [2000a], if chain propagation is viewed as being proportional to production of O<sub>3</sub>.

[33] As shown in Figures 1g, 1h, and 2c, the ratios ΔO<sub>3</sub>/ΔHNO<sub>3</sub>, ΔO<sub>3</sub>/ΔNO<sub>z</sub>, and ΔO<sub>3</sub>/ΔNO<sub>y</sub> show a good correlation with NO<sub>x</sub>-VOC sensitivity for O<sub>3</sub> below 250 ppb. The values of these ratios along the sensitivity transition still vary from low to high emissions. The correlation between these ratios and NO<sub>x</sub>-VOC sensitivity becomes substantially worse for O<sub>3</sub> higher than 250 ppb, suggesting that these ratios may behave differently in highly polluted environments. The test ratio ΔO<sub>3</sub>/S<sub>H</sub> (Figure 1i) also remains nearly constant along the sensitivity transition for O<sub>3</sub> below 250 ppb, and decreases with increasing emissions for higher O<sub>3</sub>. This ratio can also be interpreted as net ozone production efficiency per primary radical production (OPE<sub>R</sub>) and compared with previous results from Daum *et al.* [2000]. Daum *et al.* reported OPE<sub>R</sub> from 1 to 4, higher than the values shown here. However, Daum's values represent production only and do not include loss terms.

[34] The ratios O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) are especially significant because they are directly related to the error terms in equation (6a). The relation can be expressed as follows:

$$\frac{O_3}{(2H_2O_2 + NO_z)} = \left[ \frac{S_H}{(2H_2O_2 + NO_z)} \right] \left[ \frac{O_3}{S_H} \right] \quad (7a)$$

or, in difference form

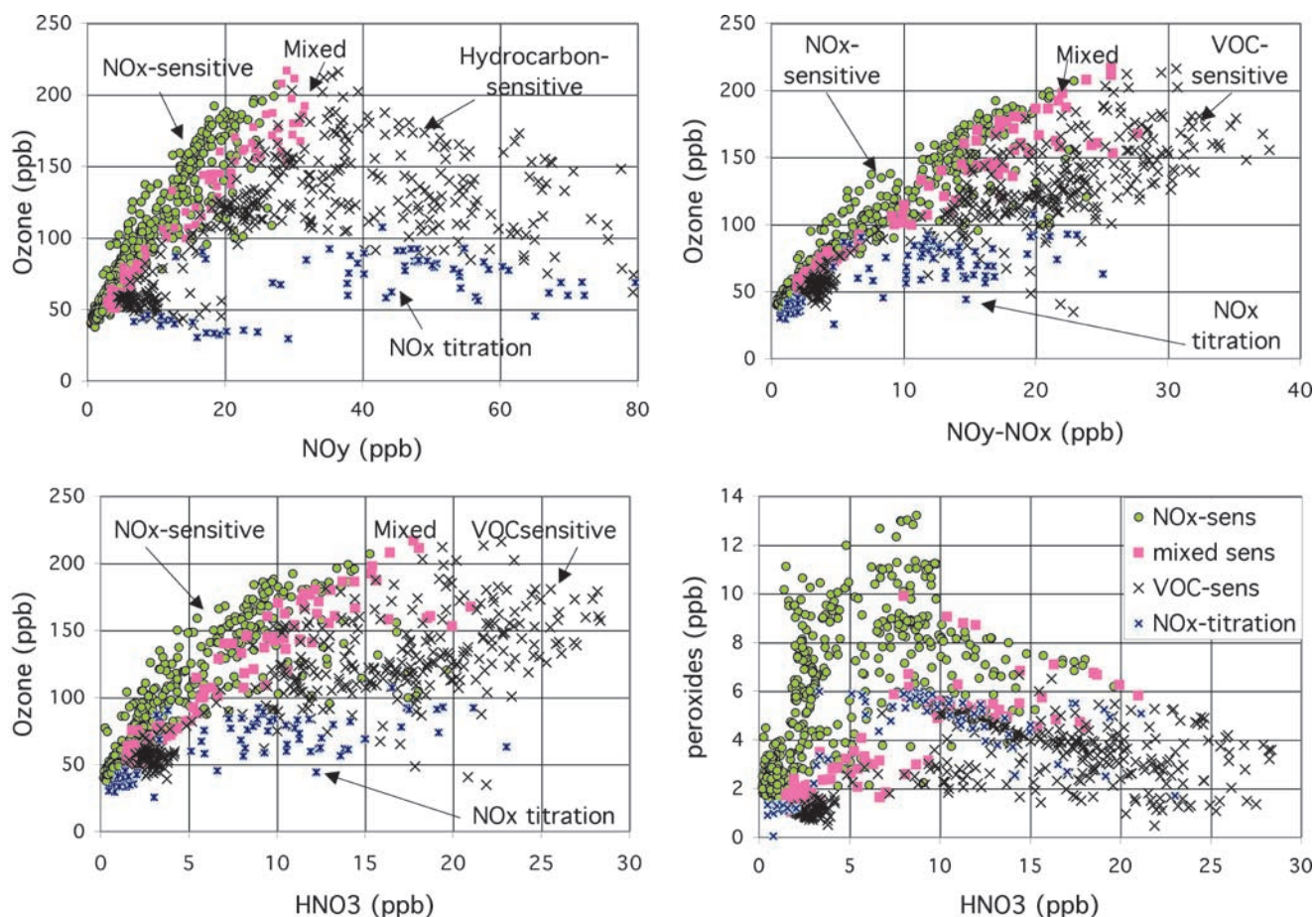
$$\frac{\Delta O_3}{\Delta(2H_2O_2 + NO_z)} = \left[ \frac{S_H}{\Delta(2H_2O_2 + NO_z)} \right] \left[ \frac{\Delta O_3}{S_H} \right] \quad (7b)$$

[35] Since H<sub>2</sub>O<sub>2</sub> and NO<sub>z</sub> represent the major radical sinks, the ratio S<sub>H</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) is determined almost entirely by the removal rate of H<sub>2</sub>O<sub>2</sub> and NO<sub>z</sub>. This is analogous to the term L<sub>N</sub>/NO<sub>z</sub> in equation (6a), which is determined by the removal rate of NO<sub>z</sub>. The terms O<sub>3</sub>/S<sub>H</sub> and ΔO<sub>3</sub>/S<sub>H</sub> appear identically in equations (6a), (6b), (7a), and (7b). Thus, the ratios O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) can be interpreted as a test for errors in the indicator ratio O<sub>3</sub>/NO<sub>z</sub>. Because they are associated with ΔO<sub>3</sub>/S<sub>H</sub>, these ratios can also be associated with ozone production efficiency per primary radical production and with chain length.

[36] As shown in Figures 1j and 2b, the ratios O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) show patterns of variation that are very similar to O<sub>3</sub>/S<sub>H</sub>. Previous results from 3-D models and measurements [Sillman *et al.*, 1998] suggested that these ratios should have a near-constant value in photochemically aged air. The results here show a broader pattern of variation. The ratio tends to decrease with increasing O<sub>3</sub> along the sensitivity transition, and also decreases (though not consistently) with increasing VOC. The range of ratio values identified by Sillman *et al.* [1998] (O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) = 6–8) are comparable with the values shown in Figure 2c for the appropriate range of O<sub>3</sub> (80–140 ppb). The difference ratio ΔO<sub>3</sub>/Δ(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) (Figures 1k and 2c) shows a pattern of variation that is comparable to ΔO<sub>3</sub>/S<sub>H</sub>.

[37] Figure 2b shows a direct comparison between the values of O<sub>3</sub>/NO<sub>z</sub>, O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) and other similar ratios along the NO<sub>x</sub>-VOC transition. As shown, these ratios all show a similar tendency to decrease as conditions range from relatively clean to highly polluted. This result is important because it suggests that O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) provides a measurement-based test for the NO<sub>x</sub>-VOC transition value associated with the indicator ratios. The NO<sub>x</sub>-VOC transition values for indicators represent model predictions and cannot be tested readily against measurements. Since O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>z</sub>) and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>z</sub>) are correlated to the indicator transition values but are not themselves correlated with NO<sub>x</sub>-VOC sensitivity, they provide an indirect test for the transition values. A similar correlation is found for the difference ratios (Figure 2c).

[38] The behavior of O<sub>3</sub>/NO<sub>z</sub> at high VOC is also influenced by the increasingly dominant role of PAN and other organic nitrates. In contrast with HNO<sub>3</sub>, the ratio O<sub>3</sub>/PAN does not correlate with NO<sub>x</sub>-VOC sensitivity (see Figure 1k) [see also Tonnesen and Dennis, 2000b]. At warm temperatures and relatively high NO<sub>x</sub> (>0.5 ppb) PAN formation reaches an approximate steady state that is proportional to O<sub>3</sub>, and the ratio O<sub>3</sub>/PAN decreases with increasing VOC [see Sillman *et al.*, 1990]. The correlation between O<sub>3</sub>/NO<sub>z</sub> and NO<sub>x</sub>-VOC sensitivity breaks down



**Figure 3.** Correlations for (a) O<sub>3</sub> versus NO<sub>y</sub>, (b) O<sub>3</sub> versus NO<sub>2</sub>, (c) O<sub>3</sub> versus HNO<sub>3</sub>, and (d) total peroxides versus HNO<sub>3</sub> (all in ppb) at 1500–1600 from the 3-D simulations listed in Table 1. Each location is classified as NO<sub>x</sub>-sensitive (green circles), VOC-sensitive (crosses), mixed or with near-zero sensitivity (lavender squares), and dominated by NO<sub>x</sub> titration (blue asterisks) based on definitions in the text.

when PAN and other organic nitrates become the dominant component of NO<sub>z</sub>.

[39] In summary, several of the ratios previously identified by *Sillman* [1995] as NO<sub>x</sub>-VOC indicators (especially O<sub>3</sub>/NO<sub>z</sub> and O<sub>3</sub>/HNO<sub>3</sub>) appear to show systematic variations in behavior as conditions vary from relatively clean to highly polluted. These variations can be understood conceptually if ratios are viewed as imperfect representations of the theoretical ratio S<sub>H</sub>/L<sub>N</sub>, which appears to correlate closely with NO<sub>x</sub>-VOC sensitivity both in these 0-D calculations and in previous work by *Kleinman et al.* [1997]. Variations in the behavior of indicator ratios in models is also associated with predicted variations in the ratios O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + NO<sub>2</sub>) and O<sub>3</sub>/(2H<sub>2</sub>O<sub>2</sub> + 2ROOH + NO<sub>2</sub>), which might be compared with ambient measurements. The next section shows equivalent results from 3-D models.

## 5. Results From 3-D Calculations

[40] Figure 3 shows O<sub>3</sub> versus NO<sub>z</sub> (afternoon values at the surface) from a composite of 3-D simulations (see summary in Table 1). The simulations include a series of regional-scale simulations in the eastern U.S. (with horizontal extent 1000 km or larger and horizontal resolution 20

× 20 km or 5 × 5 km) and for the San Joaquin valley (California), and urban-scale simulations (with horizontal extent 150–250 km) for Atlanta and Los Angeles. The figure includes results from a modified form of the simulation from *Lu and Chang* [1998] [see *Sillman et al.*, 2001], which included results that contradicted the original analysis of indicator ratios by *Sillman* [1995]. Results represent afternoon hours (1500–1600 LST).

[41] The figure also shows predicted NO<sub>x</sub>-VOC sensitivity for each location, based on differences between O<sub>3</sub> in each model base case and in equivalent scenarios with 25% or 35% reductions in anthropogenic VOC and in NO<sub>x</sub>. Locations have been classified according to the following definitions:

1. *NO<sub>x</sub>-sensitive*: O<sub>3</sub> in the scenario with reduced NO<sub>x</sub> is lower than O<sub>3</sub> in both the base case and in the scenario with reduced VOC at the specified location and hour by at least 5 ppb.
2. *VOC-sensitive*: O<sub>3</sub> in the scenario with reduced VOC is lower than O<sub>3</sub> in both the base case and in the scenario with reduced NO<sub>x</sub> by at least 5 ppb.
3. *Mixed*: The scenarios with reduced NO<sub>x</sub> and reduced VOC have O<sub>3</sub> within 5 ppb of each other, and both have O<sub>3</sub> lower than in the base case by at least 5 ppb.



4. *NO<sub>x</sub>-titration*: O<sub>3</sub> in the scenario with reduced NO<sub>x</sub> is larger than O<sub>3</sub> in the base case by at least 5 ppb, and O<sub>3</sub> in the simulation with reduced VOC is not lower by 5 ppb or more relative to the base case.

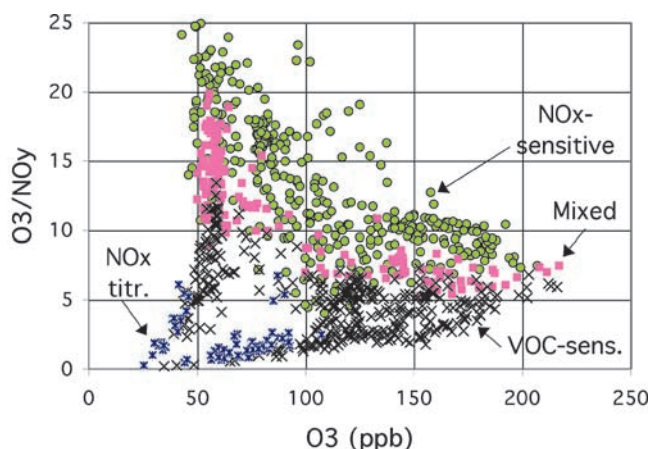
[42] All other locations are viewed as insensitive to NO<sub>x</sub> and VOC in the context of the model domain. These typically represent locations with O<sub>3</sub> dominated by transport from outside the model boundary rather than calculated photochemical production.

[43] The locations dominated by NO<sub>x</sub>-titration are usually near large sources of NO. These locations typically have relatively low O<sub>3</sub>, and O<sub>3</sub> has been affected primarily by the reaction  $O_3 + NO \rightarrow NO_2$  in the presence of directly emitted NO rather than by ozone production chemistry. Indicator ratios in these locations can be misleading because the indicator ratios involve photochemical reaction products associated with ozone formation rather than removal. Evaluation of indicator ratios will be based on the ability to distinguish between NO<sub>x</sub>-sensitive and VOC-sensitive ozone production, ignoring the locations dominated by NO<sub>x</sub>-titration.

[44] As shown in Figure 3, NO<sub>x</sub>-sensitive and VOC-sensitive locations are associated with a different range of values in the correlation plots for O<sub>3</sub> versus NO<sub>y</sub> and other ratios. The mixed locations occupy intermediate values between the NO<sub>x</sub>-sensitive and VOC-sensitive locations. Results for peroxides versus HNO<sub>3</sub> are especially useful for identifying NO<sub>x</sub>-VOC sensitivity because there is a large difference between model values associated with NO<sub>x</sub>-sensitive and VOC-sensitive locations. For O<sub>3</sub>/NO<sub>y</sub>, O<sub>3</sub>/NO<sub>z</sub> and O<sub>3</sub>/HNO<sub>3</sub> the difference between NO<sub>x</sub>-sensitive and VOC-sensitive locations is smaller and therefore more likely to be affected by model uncertainties. Also, locations dominated by NO<sub>x</sub> titration cannot be distinguished from other locations based on peroxides versus HNO<sub>3</sub>, because the process of NO<sub>x</sub> titration has little impact on peroxides or HNO<sub>3</sub>. NO<sub>x</sub>-titration locations are clearly identifiable in the plot of O<sub>3</sub> versus NO<sub>y</sub>.

[45] Previously *Sillman* [1995] associated NO<sub>x</sub>-sensitive and VOC-sensitive conditions with high and low values of indicator ratios (O<sub>3</sub>/NO<sub>y</sub>, etc.) and identified transition values that separated NO<sub>x</sub>-sensitive and VOC-sensitive locations (O<sub>3</sub>/NO<sub>y</sub> = 6–8, O<sub>3</sub>/NO<sub>z</sub> = 8–10, O<sub>3</sub>/HNO<sub>3</sub> = 12–15). A careful examination of Figure 3 shows that the NO<sub>x</sub>-VOC transition does not correspond exactly to these transition values, and that the transition values of indicator ratios decreases with increasing O<sub>3</sub>. The change in indicator transition values is especially noticeable for O<sub>3</sub> less than 80 ppb (with NO<sub>x</sub>-VOC transitions at O<sub>3</sub>/NO<sub>y</sub> = 11–15, O<sub>3</sub>/NO<sub>z</sub> = 15–20). For O<sub>3</sub> greater than 100 ppb the NO<sub>x</sub>-VOC transition corresponds to the ratio values previously identified by *Sillman* [1995]. By contrast, NO<sub>x</sub>-VOC transition for ratios involving peroxides (H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> = 0.2–0.3; total peroxides/HNO<sub>3</sub> = 0.3–0.45) do not appear to vary for the range of clean and polluted conditions shown here.

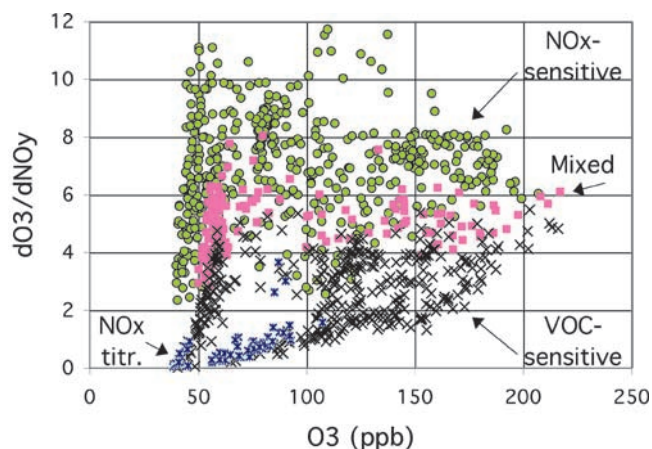
[46] Figure 4 shows the same results in different format, with the indicator ratio O<sub>3</sub>/NO<sub>y</sub> plotted versus O<sub>3</sub> for NO<sub>x</sub>-sensitive and VOC-sensitive locations. This variation in transition ratios in Figures 3 and 4 is broadly consistent with the indicator transition values identified from the 0-D simulations (see Figure 2).



**Figure 4.** O<sub>3</sub>/NO<sub>y</sub> versus O<sub>3</sub> in the 3-D simulations listed in Table 1, with locations classified as NO<sub>x</sub>-sensitive (green circles), VOC-sensitive (crosses), mixed or with near-zero sensitivity (lavender squares), and dominated by NO<sub>x</sub> titration (blue asterisks).

[47] From Figure 3 it appears that the NO<sub>x</sub>-VOC transition might fit more closely to an indicator ratio of the form  $\Delta O_3/\Delta NO_y$ , where  $\Delta O_3$  represents the difference between O<sub>3</sub> at the location in question and the model background O<sub>3</sub> (40 ppb). Indicator ratios of the form  $\Delta O_3/\Delta NO_y$  would also be advantageous for evaluating NO<sub>x</sub>-VOC sensitivity in specific urban areas that include significant transport of O<sub>3</sub> from upwind. As pointed out by *Dommen et al.* [1999] measured indicator values in plumes from relatively small cities reflect NO<sub>x</sub>-VOC sensitivity associated with region-wide production of O<sub>3</sub> rather than the impact of emissions in the individual urban center. Indicators with the form  $\Delta O_3/\Delta NO_y$  could provide information about individual metropolitan area emissions as well as region-wide conditions. However, determination of background values provides an additional uncertainty for indicators with this form.

[48] In order to identify background concentrations to be used in forming these ratios, the following formalism will be used: background concentrations will be based on the model grid with minimum NO<sub>y</sub> over the relevant portion of the model domain at the hour of interest. Here, we assume that the region of interest might be either (1) an urban plume surrounded by “background” rural concentrations; or (2) a polluted region (such as the eastern half of the U.S.) with well defined upwind conditions. We also assume that there is a specified (afternoon) hour of interest, at a time when surface conditions reflect photochemical production of O<sub>3</sub> in a well-mixed layer. Background concentrations will be set equal to conditions in the model grid with minimum NO<sub>y</sub> at the same hour and for a specified altitude. The location with minimum NO<sub>y</sub> is selected based on a horizontal domain that includes the rural background region surrounding the plume or polluted region. As suggested by *Blanchard and Stoeckenius* [2001], the model impact of reduced NO<sub>x</sub> and VOC predicted for the location with minimum NO<sub>x</sub> is subtracted from the model impact of reduced NO<sub>x</sub> and VOC at other locations, so that the analysis will not include sensitivity to emissions upwind of the boundary point. Although somewhat arbitrary, this



**Figure 5.**  $\Delta O_3/\Delta NO_y$  versus  $O_3$  in the 3-D simulations listed in Table 1, with locations classified as NO<sub>x</sub>-sensitive (green circles), VOC-sensitive (crosses), mixed or with near-zero sensitivity (lavender squares), and dominated by NO<sub>x</sub> titration (blue asterisks).  $\Delta O_3$  and  $\Delta NO_y$  refer to differences relative to background values.

method is advantageous because the same ratios and equivalent background values can easily be determined from a network of measurements that includes samples within an urban plume and the surrounding rural region. It also avoids the use of morning values to determine background, because these values are influenced by the nocturnal boundary layer and uncertain rates of entrainment from aloft. We have also avoided using the O<sub>3</sub>-NO<sub>2</sub> slope, which is often reported for measured data sets, because the value of this slope is often unclear and may depend on the precise location of the measurements.

[49] The  $\Delta O_3/\Delta NO_y$  and other ratios defined in this manner provide specific values at each model location (or at each measurement site) other than the site identified as the rural background. In the subsequent analysis we will use the above formalism to define background values for urban and regional subdomains of individual model scenarios. The model indicator ratios based on these background values will be compared with the predicted impact of reduced NO<sub>x</sub> and VOC emissions within the equivalent subdomain.

[50] As shown in Figure 5, the ratio  $\Delta O_3/\Delta NO_y$  is strongly correlated with NO<sub>x</sub>-VOC sensitivity for the group of 3-D simulations included here. There is some overlap between NO<sub>x</sub>-sensitive and VOC-sensitive locations, with NO<sub>x</sub>-sensitive values as low as 3 and VOC-sensitive values as high as 5. The NO<sub>x</sub>-VOC transition value appears to be similar for locations with high and low O<sub>3</sub>, although the value of  $\Delta O_3/\Delta NO_y$  for O<sub>3</sub> below 60 ppb is especially sensitive to the assumed background value.

[51] The information in Figure 5 can be conveniently summarized and evaluated by reporting the distribution of ratio values associated with NO<sub>x</sub>-sensitive and VOC-sensitive locations. In Figure 5 the indicator values for VOC-sensitive locations are lower than values for NO<sub>x</sub>-sensitive locations, and there is little overlap between the ranges of values. Locations with mixed sensitivity typically have intermediate indicator values. As was done in *Sillman et al.* [1998] the indicator results for individual simulations are summarized by recording the 5th, 50th and 95th percentile

indicator values for NO<sub>x</sub>-sensitive and VOC-sensitive locations. A NO<sub>x</sub>-VOC indicator is successful if the 95th percentile value for VOC-sensitive locations is equal to or less than the 5th percentile value for NO<sub>x</sub>-sensitive locations, and if the transition values remain similar in models for different locations and assumptions. These 95th percentile and 5th percentile values identify the transition from NO<sub>x</sub>-sensitive to VOC-sensitive conditions.

[52] As shown in Table 2, the ratios  $\Delta O_3/\Delta NO_y$ ,  $\Delta O_3/\Delta NO_z$  and  $\Delta O_3/\Delta HNO_3$  (where HNO<sub>3</sub> is interpreted to include aerosol nitrate) are potentially useful as NO<sub>x</sub>-VOC indicators. These ratios are correlated with NO<sub>x</sub>-VOC sensitivity both at the regional scale and within urban subsets of model domains, with  $\Delta O_3$ , etc. relative to the appropriate background for the subdomain. Median values of the ratios for NO<sub>x</sub>-sensitive locations are often twice as high as median values for VOC-sensitive locations, and the 95th percentile VOC-sensitive value and 5th percentile NO<sub>x</sub>-sensitive value are generally comparable. Ratio values at the NO<sub>x</sub>-VOC transition are:  $\Delta O_3/\Delta NO_y = 4-6$ ;  $\Delta O_3/\Delta NO_z = 5-7$ ; and  $\Delta O_3/\Delta HNO_3 = 8-10$ . It is noteworthy that range of values from the model for San Joaquin reported by *Lu and Chang* [1998] are generally comparable to the others. Results for O<sub>3</sub>/NO<sub>z</sub> from this model differed sharply from the indicator results in *Sillman* [1995], probably because the scenario reported by San Joaquin had different boundary conditions [see *Sillman et al.*, 2001]. When ratios such as  $\Delta O_3/\Delta NO_y$  along with the described procedure for identifying background values are used, model boundary conditions are less important.

[53] Although the ratios  $\Delta O_3/\Delta NO_y$ , etc. effectively separate NO<sub>x</sub>-sensitive and VOC-sensitive conditions for many of the 3-D simulations, there were some situations with ambiguous results. The ratios  $\Delta O_3/\Delta NO_z$  and  $\Delta O_3/\Delta HNO_3$  performed poorly in the models for Los Angeles, San Joaquin and in one Nashville case. In the ambiguous cases, the 5th percentile ratio value for NO<sub>x</sub>-sensitive locations is comparable to the median value for VOC-sensitive locations, suggesting significant overlap of values between NO<sub>x</sub>-sensitive and VOC-sensitive conditions. Performance is significantly better for  $\Delta O_3/\Delta NO_y$  in these locations. Indicator results are especially poor for the model for Los Angeles reported by *Chock et al.* [1999]. Transition values can also be shifted by 20% in response to changes in dry deposition rates.

[54] Table 2 also includes results for the original indicator ratios (O<sub>3</sub>/NO<sub>y</sub>, H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub>, etc.) from *Sillman* [1995]. These ratios generally show a larger separation between NO<sub>x</sub>-sensitive and VOC-sensitive locations than the ratios with the form  $\Delta O_3/\Delta NO_y$ , etc. They also avoid the uncertainty associated with background values for  $\Delta O_3/\Delta NO_y$ . As described above, the NO<sub>x</sub>-VOC transition for the ratios O<sub>3</sub>/NO<sub>y</sub>, O<sub>3</sub>/NO<sub>z</sub> and O<sub>3</sub>/HNO<sub>3</sub> are much higher in the models for San Joaquin and in the model for Los Angeles reported by *Chock et al.* than in other cases.

[55] The ratio H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> and similar ratios involving summed peroxides generally show the strongest correlation with NO<sub>x</sub>-VOC sensitivity and the smallest variation among individual models. However, the difference ratio relative to background ( $\Delta H_2O_2/\Delta HNO_3$ ) does not correlate with NO<sub>x</sub>-VOC sensitivity. Typically, H<sub>2</sub>O<sub>2</sub> is lower inside an urban plume relative to the surrounding rural area [*Weinstein-*

**Table 2.** Distribution of Indicator Values for NO<sub>x</sub>- and VOC-Sensitive Locations in 3-D Simulations

| Indicator                               | VOC-Sensitive Locations |                 |                           | NO <sub>x</sub> -Sensitive Locations |                 |                 |
|---|-------------------------|-----------------|---------------------------|--------------------------------------|-----------------|-----------------|
|   | 5th Percentile          | 50th Percentile | 95th Percentile           | 5th Percentile                       | 50th Percentile | 95th Percentile |
|   |                         |                 | $\Delta O_3/\Delta NO_y$  |                                      |                 |                 |
| Nashville, full domain                  | 0.4                     | 2.9             | 3.6                       | 3.4                                  | 4.7             | 5.3             |
| Nashville, high deposition <sup>a</sup> | 0.3                     | 3.5             | 4.3                       | 4.6                                  | 6.7             | 7.7             |
| Nashville, urban subdomain              | 1.0                     | 2.5             | 3.6                       | 4.2                                  | 5.6             | 12.4            |
| Northeast, full domain                  | 2.8                     | 4.0             | 4.8                       | 5.7                                  | 7.7             | 9.0             |
| Northeast, urban subdomain              | 1.8                     | 3.8             | 5.0                       | 5.0                                  | 6.5             | 7.2             |
| Lake Michigan                           | 2.4                     | 3.4             | 4.3                       | 3.5                                  | 6.2             | 7.6             |
| Atlanta                                 | 1.6                     | 2.8             | 4.5                       | 4.5                                  | 6.9             | 11.1            |
| San Joaquin (Sillman)                   | 0.6                     | 2.8             | 4.9                       | 4.4                                  | 7.5             | 11.4            |
| San Joaquin (Lu and Chang)              | 1.3                     | 3.2             | 4.2                       | 3.6                                  | 8.4             | 32.4            |
| Los Angeles (Godowitch)                 | 0.9                     | 2.3             | 4.6                       | 5.2                                  | 7.9             | 11.9            |
| Los Angeles (Chock)                     | 0.6                     | 3.7             | 6.0                       | 3.5                                  | 8.6             | 13.7            |
|   |                         |                 | $\Delta O_3/\Delta NO_2$  |                                      |                 |                 |
| Nashville, full domain                  | 1.2                     | 3.6             | 5.0                       | 3.8                                  | 5.3             | 6.1             |
| Nashville, high deposition <sup>a</sup> | 1.6                     | 4.8             | 6.6                       | 5.3                                  | 8.1             | 9.1             |
| Nashville, urban subdomain              | 1.9                     | 4.0             | 5.0                       | 5.1                                  | 7.1             | 15.4            |
| Northeast, full domain                  | 4.7                     | 5.0             | 5.6                       | 6.7                                  | 8.5             | 9.9             |
| Northeast, urban subdomain              | 2.8                     | 4.4             | 5.4                       | 5.7                                  | 7.4             | 8.3             |
| Lake Michigan                           | 3.0                     | 4.0             | 5.2                       | 5.7                                  | 7.3             | 8.5             |
| Atlanta                                 | 3.6                     | 4.5             | 5.9                       | 6.1                                  | 9.2             | 14.6            |
| San Joaquin (Sillman)                   | 2.4                     | 6.1             | 8.1                       | 6.0                                  | 9.9             | 15.2            |
| San Joaquin (Lu and Chang)              | 3.5                     | 6.4             | 6.6                       | 5.0                                  | 11.6            | 31.4            |
| Los Angeles (Godowitch)                 | 3.0                     | 4.5             | 6.3                       | 6.2                                  | 9.0             | 15.6            |
| Los Angeles (Chock)                     | 4.2                     | 6.7             | 10.0                      | 8.3                                  | 10.9            | 16.6            |
|   |                         |                 | $\Delta O_3/\Delta HNO_3$ |                                      |                 |                 |
| Nashville, full domain                  | 1.4                     | 4.2             | 6.5                       | 4.6                                  | 6.7             | 8.2             |
| Nashville, high deposition <sup>a</sup> | 1.8                     | 5.9             | 9.4                       | 7.1                                  | 11.5            | 14.1            |
| Nashville, urban subdomain              | 2.2                     | 5.1             | 7.9                       | 8.0                                  | 12.8            | 40.3            |
| Northeast, full domain                  | 7.3                     | 8.4             | 10.7                      | 11.0                                 | 15.6            | 23.9            |
| Northeast, urban subdomain              | 4.9                     | 8.4             | 11.4                      | 13.1                                 | 27.3            | 235.7           |
| Lake Michigan                           | 3.6                     | 4.8             | 7.6                       | 8.8                                  | 12.5            | 18.2            |
| Atlanta                                 | 4.1                     | 5.3             | 7.3                       | 7.3                                  | 13.1            | 20.5            |
| San Joaquin (Sillman)                   | 2.7                     | 7.4             | 11.0                      | 9.4                                  | 14.7            | 29.6            |
| Los Angeles (Godowitch)                 | 3.5                     | 5.5             | 8.9                       | 9.1                                  | 14.6            | 31.2            |
| Los Angeles (Chock)                     | 6.2                     | 10.1            | 25.3                      | 13.1                                 | 18.3            | 42.4            |
|   |                         |                 | $O_3/NO_y$                |                                      |                 |                 |
| Nashville                               | 2.2                     | 3.6             | 5.6                       | 7.1                                  | 10.8            | 13.             |
| Nashville, high dep. <sup>a</sup>       | 2.6                     | 5.3             | 7.2                       | 8.7                                  | 13.9            | 17.             |
| Northeast corridor                      | 5.0                     | 5.4             | 6.5                       | 8.2                                  | 12.7            | 18.             |
| Lake Michigan                           | 3.5                     | 5.2             | 6.6                       | 7.2                                  | 11.7            | 16.             |
| Atlanta                                 | 3.6                     | 5.1             | 7.2                       | 8.1                                  | 14.3            | 27.             |
| Los Angeles (Godowitch)                 | 1.4                     | 3.2             | 6.0                       | 7.2                                  | 10.4            | 20.             |
| Los Angeles (Chock)                     | 1.0                     | 5.4             | 9.1                       | 5.9                                  | 18.             | 24.             |
| San Joaquin (Sillman)                   | 3.0                     | 7.3             | 11.6                      | 15.                                  | 26.             | 56.             |
| San Joaquin (Chang)                     | 5.4                     | 7.7             | 12.3                      | 12.5                                 | 22.             | 52.             |
|   |                         |                 | $O_3/NO_2$                |                                      |                 |                 |
| Nashville                               | 5.3                     | 6.3             | 7.3                       | 8.4                                  | 12.4            | 13.             |
| Nashville, high dep. <sup>a</sup>       | 7.0                     | 8.8             | 9.9                       | 10.5                                 | 16.8            | 20.             |
| Northeast corridor                      | 6.7                     | 7.1             | 7.7                       | 9.7                                  | 14.4            | 20.             |
| Lake Michigan                           | 4.5                     | 5.8             | 8.1                       | 11.2                                 | 15.3            | 18.             |
| Atlanta                                 | 7.1                     | 8.6             | 10.0                      | 11.9                                 | 17.7            | 33.             |
| Los Angeles (Godowitch)                 | 4.6                     | 6.2             | 8.4                       | 8.7                                  | 12.0            | 25.             |
| Los Angeles (Chock)                     | 6.2                     | 10.             | 15                        | 14.                                  | 23.             | 31.             |
| San Joaquin (Sillman)                   | 11.                     | 16.             | 19.                       | 21.                                  | 35.             | 73.             |
| San Joaquin (Chang)                     | 14.                     | 16.             | 20.                       | 17.                                  | 32.             | 68.             |
|   |                         |                 | $O_3/HNO_3$               |                                      |                 |                 |
| Nashville                               | 6.                      | 8.              | 9.                        | 12.                                  | 19.             | 21.             |
| Nashville, high dep. <sup>a</sup>       | 8.                      | 11.             | 14.                       | 15.                                  | 25.             | 29.             |
| Northeast corridor                      | 10.                     | 11.             | 16.                       | 17.                                  | 27.             | 42.             |
| Lake Michigan                           | 5.                      | 7.              | 11.                       | 16.                                  | 28.             | 39.             |
| Atlanta                                 | 8.                      | 10.             | 13.                       | 15.                                  | 25.             | 54.             |
| Los Angeles (Godowitch)                 | 5.                      | 8.              | 12.                       | 13.                                  | 19.             | 50.             |
| Los Angeles (Chock)                     | 9.                      | 15.             | 40.                       | 23.                                  | 39.             | 80.             |
| San Joaquin (Sillman)                   | 14.                     | 20.             | 28.                       | 27.                                  | 54.             | 155.            |

**Table 2.** (continued)

| Indicator                            | VOC-Sensitive Locations                                   |                 |                 | NO <sub>x</sub> -Sensitive Locations |                 |                 |
|--------------------------------------|---|-----------------|-----------------|--------------------------------------|-----------------|-----------------|
|                                      | 5th Percentile  | 50th Percentile | 95th Percentile | 5th Percentile                       | 50th Percentile | 95th Percentile |
|                                      | <i>H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub></i>         |                 |                 |                                      |                 |                 |
| Nashville                            | 0.11  | 0.18            | 0.23            | 0.30                                 | 0.64            | 0.82            |
| Nashville, high dep. <sup>a</sup>    | 0.09  | 0.15            | 0.21            | 0.26                                 | 0.55            | 0.70            |
| Northeast corridor                   | 0.22  | 0.33            | 0.40            | 0.68                                 | 1.58            | 3.1             |
| Lake Michigan                        | 0.03  | 0.13            | 0.24            | 0.37                                 | 1.15            | 1.8             |
|                                      | <i>(H<sub>2</sub>O<sub>2</sub>+ ROOH)/HNO<sub>3</sub></i> |                 |                 |                                      |                 |                 |
| Nashville                            | 0.18  | 0.23            | 0.31            | 0.39                                 | 0.73            | 1.01            |
| Nashville, high dep. <sup>a</sup>    | 0.22  | 0.28            | 0.37            | 0.54                                 | 1.14            | 1.48            |
| Northeast corridor                   | 0.22  | 0.39            | 0.54            | 0.93                                 | 2.0             | 4.3             |
| Lake Michigan                        | 0.06  | 0.20            | 0.32            | 0.49                                 | 1.6             | 2.6             |
| Atlanta <sup>b</sup>                 | 0.17  | 0.27            | 0.55            | 0.55                                 | 1.44            | 4.5             |
| Los Angeles (Godowitch) <sup>b</sup> | 0.06  | 0.22            | 0.41            | 0.40                                 | 0.87            | 2.8             |
| Los Angeles (Chock) <sup>b</sup>     | 0.12  | 0.33            | 1.8             | 0.80                                 | 2.2             | 4.4             |
| San Joaquin (Sillman) <sup>b</sup>   | 0.24  | 0.38            | 0.56            | 1.0                                  | 2.7             | 8.7             |
| San Joaquin (Chang) <sup>b</sup>     | 0.17  | 0.23            | 0.52            | 0.72                                 | 1.8             | 6.3             |
|                                      | <i>H<sub>2</sub>O<sub>2</sub>/NO<sub>z</sub></i>          |                 |                 |                                      |                 |                 |
| Nashville                            | 0.12  | 0.15            | 0.19            | 0.21                                 | 0.41            | 0.45            |
| Nashville, high dep. <sup>a</sup>    | 0.10  | 0.13            | 0.17            | 0.18                                 | 0.37            | 0.41            |
| Northeast corridor                   | 0.14  | 0.20            | 0.21            | 0.36                                 | 0.80            | 1.4             |
| Lake Michigan                        | 0.03  | 0.10            | 0.19            | 0.27                                 | 0.60            | 0.87            |
|                                      | <i>(H<sub>2</sub>O<sub>2</sub>+ ROOH)/NO<sub>z</sub></i>  |                 |                 |                                      |                 |                 |
| Nashville                            | 0.13  | 0.20            | 0.27            | 0.29                                 | 0.59            | 0.85            |
| Nashville, high dep. <sup>a</sup>    | 0.15  | 0.22            | 0.30            | 0.35                                 | 0.79            | 1.08            |
| Northeast corridor                   | 0.14  | 0.24            | 0.28            | 0.49                                 | 1.10            | 1.91            |
| Lake Michigan                        | 0.05  | 0.15            | 0.26            | 0.35                                 | 0.81            | 1.22            |
| Atlanta <sup>b</sup>                 | 0.14  | 0.23            | 0.44            | 0.42                                 | 1.04            | 2.8             |
| Los Angeles (Godowitch) <sup>b</sup> | 0.05  | 0.17            | 0.33            | 0.26                                 | 0.55            | 1.4             |
|                                      | <i>H<sub>2</sub>O<sub>2</sub>/NO<sub>y</sub></i>          |                 |                 |                                      |                 |                 |
| Nashville                            | 0.05  | 0.09            | 0.12            | 0.17                                 | 0.36            | 0.42            |
| Nashville, high dep. <sup>a</sup>    | 0.04  | 0.07            | 0.10            | 0.14                                 | 0.32            | 0.40            |
| Northeast corridor                   | 0.11  | 0.14            | 0.17            | 0.30                                 | 0.68            | 1.3             |
| Lake Michigan                        | 0.02  | 0.09            | 0.16            | 0.22                                 | 0.43            | 0.78            |
|                                      | <i>(H<sub>2</sub>O<sub>2</sub>+ ROOH)/NO<sub>y</sub></i>  |                 |                 |                                      |                 |                 |
| Nashville                            | 0.08  | 0.12            | 0.18            | 0.24                                 | 0.52            | 0.98            |
| Nashville, high dep. <sup>a</sup>    | 0.08  | 0.12            | 0.17            | 0.28                                 | 0.68            | 1.15            |
| Northeast corridor                   | 0.11  | 0.19            | 0.22            | 0.38                                 | 0.9             | 1.72            |
| Lake Michigan                        | 0.04  | 0.12            | 0.25            | 0.28                                 | 0.58            | 1.09            |
| Atlanta <sup>b</sup>                 | 0.10  | 0.12            | 0.19            | 0.29                                 | 0.83            | 2.2             |
| Los Angeles (Godowitch) <sup>b</sup> | 0.02  | 0.08            | 0.19            | 0.22                                 | 0.47            | 1.2             |

The table shows 5th, 50th, and 95th percentile indicator values (with percentile ordering by indicator value) for VOC-sensitive locations and NO<sub>x</sub>-sensitive locations as defined in the text. The terms ΔO<sub>3</sub>, ΔNO<sub>y</sub>, ΔNO<sub>z</sub> and ΔHNO<sub>3</sub> represent the difference between values at specified locations and background values. In some cases results are shown separately for (a) the full model domain and (b) an urban subdomain. In these cases NO<sub>x</sub>-VOC sensitivity is defined relative to changed emissions within the model subdomain and background values are determined relative to the same subdomain. Models are described in Table 1.

<sup>a</sup>The Nashville high-deposition scenario has dry deposition rates of 5 cm<sup>2</sup> s<sup>-1</sup> for H<sub>2</sub>O<sub>2</sub> and HNO<sub>3</sub>, as opposed to 2.5 cm<sup>2</sup> s<sup>-1</sup> in the standard scenario.

<sup>b</sup>In models with CB-4 chemistry, "H<sub>2</sub>O<sub>2</sub>" is interpreted as a surrogate for the sum of H<sub>2</sub>O<sub>2</sub> and organic peroxides.

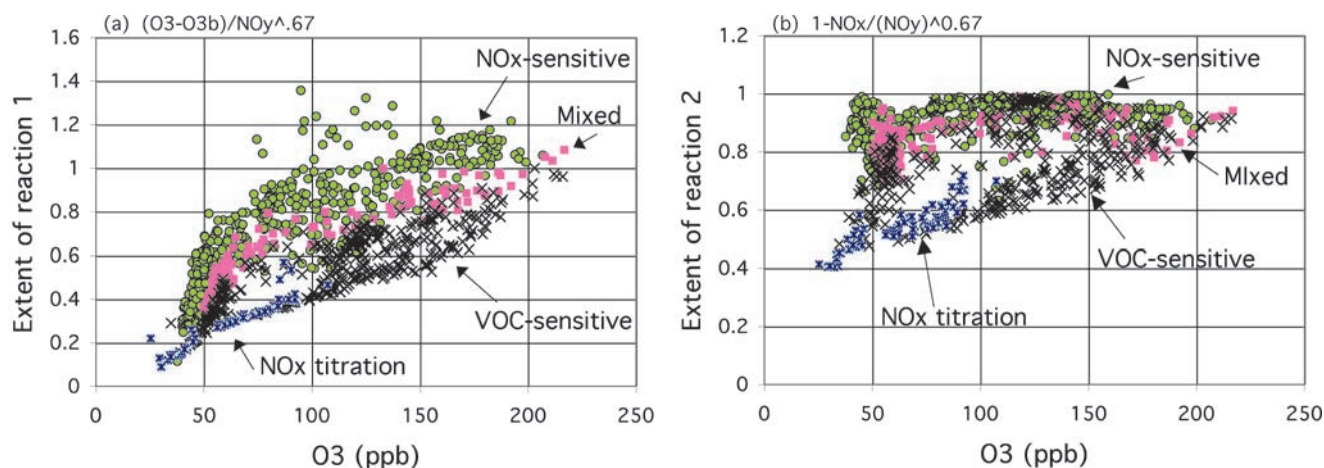
Lloyd *et al.*, 1998]. This decrease in H<sub>2</sub>O<sub>2</sub> may suggest that the urban plume has VOC-sensitive chemistry, but it may also occur because the surrounding rural area has strongly NO<sub>x</sub>-sensitive conditions. A correlation with urban NO<sub>x</sub>-VOC chemistry independent of conditions in the surrounding rural region does not appear to be possible. H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> and similar ratios correlate with NO<sub>x</sub>-VOC sensitivity on regional scales, where most of the H<sub>2</sub>O<sub>2</sub> and HNO<sub>3</sub> has been produced within the model domain of interest.

## 6. Discussion

[56] The results of the previous section should be viewed in the context of the need to identify the range of conditions for which the proposed NO<sub>x</sub>-VOC indicators are valid. The

general pattern of correlation between indicators and NO<sub>x</sub>-VOC sensitivity appears to apply for a very wide range of conditions, as illustrated in Figures 3 and 4. As pointed out by Lu and Chang [1998], the NO<sub>x</sub>-VOC transition values for some indicator ratios can vary significantly. However, this does not appear to be just a random variation among different models or locations. Rather, it is part of a systematic variation from relatively clean to highly polluted conditions, which is imperfectly captured by the indicator ratios. This variation can be explained in terms of the radical chemistry that drives the split into NO<sub>x</sub>-sensitive and VOC-sensitive conditions.

[57] The indicator transition values from Sillman [1995] and Sillman *et al.* [1998] appear to be valid for moderately polluted conditions with O<sub>3</sub> from 80 to 200 ppb. The 0-D



**Figure 6.** Extent of reaction parameters [Blanchard *et al.*, 1999] versus O<sub>3</sub> in the 3-D simulations listed in Table 1, with locations classified as NO<sub>x</sub>-sensitive (green circles), VOC-sensitive (crosses), mixed or with near-zero sensitivity (lavender squares), and dominated by NO<sub>x</sub> titration (blue asterisks). The extent parameters are (a) based on  $\Delta O_3/\Delta NO_y^{0.67}$  (equation (8a)) and (b) based on  $(NO_x/NO_y)^{0.67}$  (equation (8b)).

calculations shown here suggest that some indicator ratios (O<sub>3</sub>/NO<sub>z</sub>) may not be applicable for more highly polluted conditions, and others may have different NO<sub>x</sub>-VOC transition values. This finding is especially important in connection to investigations of the chemistry of ozone formation in Mexico City. Sosa *et al.* [2000] reported that the ratio O<sub>3</sub>/NO<sub>z</sub> did not correlate well with NO<sub>x</sub>-VOC sensitivity in 3-D models for Mexico City.

[58] From the calculations shown here, the indicator ratios involving peroxides appear to be more robust than the ratios involving O<sub>3</sub>. Lu and Chang [1998] and Chock *et al.* [1999] reported transition values for H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub> that differed from previously reported results.

[59] It may be possible to derive more robust ratios by using the form  $\Delta O_3/\Delta NO_z$  rather than the simpler O<sub>3</sub>/NO<sub>z</sub>. Indicator ratios of this form are more consistent with the pattern of variation among NO<sub>x</sub>-sensitive and VOC-sensitive locations shown in Figure 3. However, this would introduce additional uncertainty concerning the value of the background concentrations, which are needed to determine  $\Delta O_3$ . In addition, ratios with this form still show significant variation and in models for different locations. It is perhaps more useful to interpret measured indicator ratios by directly comparing correlations with the broad pattern of NO<sub>x</sub>-VOC variation in Figure 3. A direct comparison of correlation plots would also help to establish the validity of using specific sets of measurements in combination with these theoretical results.

[60] It is useful to compare indicator ratios described here with the “extent-of-reaction” parameters developed by Johnson *et al.* [1990] and Blanchard *et al.* [1999; Blanchard and Stoeckenius, 2001] [see also Olzyna *et al.*, 1994; T. Chang *et al.*, 1997]. Extent-of-reaction parameters were derived from smog chamber experiments, and are based on the finding that ozone in photochemically aged air is sensitive to NO<sub>x</sub> while ozone in air with relatively unprocessed emissions is sensitive to VOC. Although the rationale is different, the extent-of-reaction parameters are often similar to the ratio  $\Delta O_3/\Delta NO_y$  used here. Blanchard *et al.* [1999] developed the following parameters for extent of

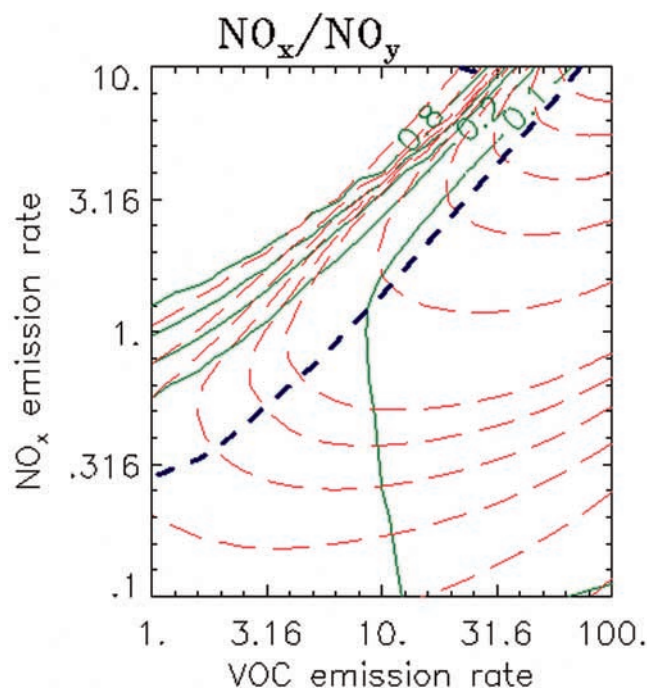
reaction (modified here to account for dry deposition of NO<sub>y</sub>):

$$\text{Extent}_1 = \frac{1.1O_3 - O_3b - NO + 1.23 NO_y}{22NO_y^{0.67}} \quad (8a)$$

$$\text{Extent}_2 = \left[ 1 - \frac{NO_x}{1.3NO_y} \right]^{0.67} \quad (8b)$$

[61] Figure 6 shows how the extent parameters vary for NO<sub>x</sub>-sensitive and VOC-sensitive locations for the ensemble of models used here. The first extent parameter (based on  $\Delta O_3/\Delta NO_y^{0.67}$ ) effectively separates NO<sub>x</sub>-sensitive from VOC-sensitive locations, but the NO<sub>x</sub>-VOC transition tends to vary from low to high O<sub>3</sub>. A better correlation with NO<sub>x</sub>-VOC sensitivity would be obtained using the original extent parameter developed by Johnson *et al.* [1990] (based on  $\Delta O_3/\Delta NO_y$ ), although the original parameter would need to be modified to account for removal of NO<sub>y</sub> by dry deposition. Blanchard *et al.* [1999] described how the 0.67 exponent was added based on results from 0-D calculations and smog chamber experiments. Based on the isopleths in Figure 1h and 2b, it is possible that the form  $\Delta O_3/\Delta NO_y^{0.67}$  may be more appropriate for highly polluted environments. The NO<sub>x</sub>-VOC transition in the isopleth plots (see Figure 2b) correlates more closely to  $\Delta O_3/\Delta NO_y^{0.67}$  than to  $\Delta O_3/\Delta NO_y$  for highly polluted conditions (O<sub>3</sub> > 150 ppb).

[62] The second extent parameter (based on  $1-NO_x/NO_y$ ) is loosely correlated with NO<sub>x</sub>-VOC sensitivity but there are a large number of exceptions, with VOC-sensitive chemistry and extent of reaction close to 1. A similar analysis was done by Blanchard and Stoeckenius [2001]. We also found poor performance for a third extent parameter, reported by Blanchard *et al.* [1999], based on O<sub>3</sub> and NO<sub>x</sub>. The poor correlation for this parameter may be related to a flaw in the extent-of-reaction concept. Figure 7 shows isopleths for NO<sub>x</sub>/NO<sub>y</sub> in comparison with O<sub>3</sub> from the 0-D calculations (equivalent to Figure 1). As shown in the figure, high NO<sub>x</sub>/



**Figure 7.** Isopleths for  $\text{NO}_x/\text{NO}_y$  (red dashed lines) as a function of  $\text{NO}_x$  and VOC emissions ( $10^{12}$  molec.  $\text{cm}^{-2}$   $\text{s}^{-1}$ ) in 0-D simulations, as in Figure 1. The solid green line represents  $\text{O}_3$ , and the blue dashed line represents the transition from VOC-sensitive to  $\text{NO}_x$ -sensitive conditions.

$\text{NO}_y$  ratios are associated with strongly VOC-sensitive chemistry in which ozone production is sharply limited by excess  $\text{NO}_x$ . The high  $\text{NO}_x/\text{NO}_y$  also corresponds to  $S_{\text{H}}/S_{\text{N}} < 1$  in Figure 1a. Low  $\text{NO}_x/\text{NO}_y$  ratios in Figure 7 correspond both to  $\text{NO}_x$ -sensitive conditions and to VOC-sensitive conditions that have relatively high rates of ozone production. In the 3-D simulations also a low value of the extent parameter (corresponding to high  $\text{NO}_x/\text{NO}_y$ ) is associated with strongly VOC-sensitive chemistry, but a high value (corresponding to low  $\text{NO}_x/\text{NO}_y$ ) can be associated with either  $\text{NO}_x$ -sensitive or VOC-sensitive conditions. The exceptional VOC-sensitive cases occur mainly in the simulation for Lake Michigan, in which a strongly VOC-sensitive plume from Chicago moves downwind and remains sensitive to VOC even after it ages. The extent-of-reaction parameters based on  $\Delta\text{O}_3/\Delta\text{NO}_y$  show a closer correlation with  $\text{NO}_x$ -VOC sensitivity in models, possibly because  $\Delta\text{O}_3/\Delta\text{NO}_y$  is associated with Kleinman's radical ratio ( $S_{\text{H}}/L_{\text{N}}$ , discussed in sections 2 and 4) rather than just the photochemical age.

## 7. Conclusion

[63] The above analysis has resulted in the following new findings concerning the use of secondary species  $\text{NO}_x$ -VOC indicators.

1. The behavior of some proposed indicator ratios tends to be different for relatively clean environments ( $\text{O}_3 < 80$  ppb), moderately polluted (100–200 ppb  $\text{O}_3$ ) and highly polluted environments ( $\text{O}_3 > 200$  ppb). This is especially true for  $\text{O}_3/\text{NO}_z$ . Previously reported results from Sillman [1995] and Sillman *et al.* [1998] were based on moderately polluted conditions.

2. The correlation with  $\text{NO}_x$ -VOC sensitivity is stronger and more consistent for ratios involving peroxides ( $\text{H}_2\text{O}_2/\text{HNO}_3$ , etc.) than for ratios such as  $\text{O}_3/\text{NO}_z$ .

3. The ratios  $\text{O}_3/(2\text{H}_2\text{O}_2 + \text{NO}_z)$  and  $\text{O}_3/(2\text{H}_2\text{O}_2 + 2\text{ROOH} + \text{NO}_z)$  should also vary from clean to highly polluted environments. These ratios provide a measurement-based test of indicator  $\text{NO}_x$ -VOC transition values.

4. Indicator ratios with the form  $\Delta\text{O}_3/\Delta\text{NO}_y$  (where  $\Delta\text{O}_3$  and  $\Delta\text{NO}_y$  represents the difference between values at a specific location and background values) are proposed for evaluating local  $\text{NO}_x$ -VOC sensitivity in urban plumes with well-defined background concentrations. This ratio is generally correlated with  $\text{O}_3$ - $\text{NO}_x$ -VOC sensitivity in models, but there is significant variation among models for different locations.

5. Variations in indicator behavior are analytically linked to variations in the ozone production efficiency per primary radical production ( $\text{O}_3/S_{\text{H}}$ ).

6. The “extent-of-reaction” parameter ( $\Delta\text{O}_3/\Delta\text{NO}_y^{0.67}$ ) also correlates with  $\text{NO}_x$ -VOC sensitivity in models, but its behavior shows a systematic variation with  $\text{O}_3$ . Extent parameters based on  $\text{NO}_x/\text{NO}_y$  do not correlate with model  $\text{NO}_x$ -VOC sensitivity.

[64] The indicator ratios discussed here are a potentially useful method for identifying  $\text{NO}_x$ -sensitive and VOC-sensitive conditions in the ambient atmosphere, but they are also subject to many uncertainties. These include: uncertain dry deposition rates, wet deposition, aerosol interactions, measurement error, and case-to-case variations. It is especially important to recognize that many of the results shown here depend on the mechanism used to represent photochemistry. Derived ratios such as the ozone production per primary radical production may depend on the detailed representation of reaction sequences of VOC, which are often highly parameterized. It is important to compare these values for different photochemical representations and compare them with measurements (including both smog chamber experiments).

[65] Perhaps the biggest weakness is the difficulty in testing the predicted indicator- $\text{NO}_x$ -VOC relationship against ambient measurements. Comparisons between measured data sets and the predicted patterns of correlation in Figure 3 are likely to be more useful in establishing the validity of indicator ratios.

## 8. Notation

|                         |  |
|-------------------------|--|
| $\text{NO}_z$           | $\text{NO}_x$ reaction products, or $\text{NO}_y$ - $\text{NO}_x$  |
| $S_{\text{H}}, Q$       | summed source of odd hydrogen radicals, including OH, $\text{HO}_2$ and $\text{RO}_2$ [referred to as Q in Kleinman <i>et al.</i> , 1997]. |
| $S_{\text{N}}$          | summed source of $\text{NO}_x$ .   |
| $L_{\text{N}}$          | summed rate of photochemical removal of $\text{NO}_x$ , also equal to the rate of production of $\text{NO}_z$ .                            |
| $\text{OPE}_{\text{N}}$ | ozone production efficiency per $\text{NO}_x$ .  |
| $\text{OPE}_{\text{R}}$ | ozone production efficiency per primary radical production.  |

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